

## CO<sub>2</sub> flux and seasonal variability in the turbidity maximum zone and surrounding area in the Changjiang River estuary\*

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**Abstract** The turbidity maximum zone (TMZ) is one of the most important regions in an estuary. However, the high concentration of suspended material makes it difficult to measure the partial pressure of CO<sub>2</sub> ( $p\text{CO}_2$ ) in these regions. Therefore, very little data is available on the  $p\text{CO}_2$  levels in TMZs. To relatively accurately evaluate the CO<sub>2</sub> flux in an example estuary, we studied the TMZ and surrounding area in the Changjiang (Yangtze) River estuary. From seasonal cruises during February, August, November 2010, and May 2012, the  $p\text{CO}_2$  in the TMZ and surrounding area was calculated from pH and total alkalinity (TA) measured in situ, from which the CO<sub>2</sub> flux was calculated. Overall, the TMZ and surrounding area acted as a source of atmosphere CO<sub>2</sub> in February and November, and as a sink in May and August. The average FCO<sub>2</sub> was -9, -16, 5, and 5 mmol/(m<sup>2</sup>·d) in May, August, November, and February, respectively. The TMZ's role as a source or sink of atmosphere CO<sub>2</sub> was quite different to the outer estuary. In the TMZ and surrounding area, suspended matter, phytoplankton, and pH were the main factors controlling the FCO<sub>2</sub>, but here the influence of temperature, salinity, and total alkalinity on the FCO<sub>2</sub> was weak. Organic carbon decomposition in suspended matter was the main reason for the region acting as a CO<sub>2</sub> source in winter, and phytoplankton production was the main reason the region was a CO<sub>2</sub> sink in summer.

**Keyword:** CO<sub>2</sub> flux; seasonal variability; turbidity maximum zone; Changjiang River estuary

### 1 INTRODUCTION

Numerous studies have shown that estuarine waters act as significant sources of CO<sub>2</sub> to the atmosphere (Zhai et al., 2007; Chen et al., 2008; Jiang et al., 2008; Guo et al., 2009; Zhai and Dai, 2009; Cai, 2011; Chen et al., 2013). Although globally, the surface area of estuaries only accounts for 4% of the continental shelf, its CO<sub>2</sub> degassing flux is as large as the CO<sub>2</sub> uptake of the entire continental shelf. Thus, both play important roles in global CO<sub>2</sub> flux (Borges, 2005; Borges et al., 2005; Cai et al., 2006; Li et al., 2007; Chen and Borges, 2009). Estimates for the global CO<sub>2</sub> degassing flux from estuaries varies, with Borges et al. (Borges, 2005; Borges et al., 2005) suggesting the flux is 0.34–0.43 pg C/a, while Cai (2011) and Chen (2012) suggested the flux is 0.25 and 0.26 pg C/a, respectively. However, there is no accepted accurate estuarine CO<sub>2</sub> degassing flux data, which indicates that our current knowledge of global

estuarine CO<sub>2</sub> efflux is largely uncertain, for several reasons.

First, a limited number of estuaries have been investigated, and clearly studies do not include all estuaries globally. Second, in general, the amount of CO<sub>2</sub> released from water varies diurnally and seasonally, however, most investigations ignore these variations. Third, within the same estuary, different regions may act as sinks or sources of CO<sub>2</sub>. In general, in large river, the upper regions of estuaries act as strong sources of CO<sub>2</sub>, and wider regions in the mid-to lower estuaries act as weak sources or sinks. However, not all investigations of surface  $p\text{CO}_2$  and

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air-water CO<sub>2</sub> flux cover the whole estuarine zone, especially the turbidity maximum zone (TMZ).

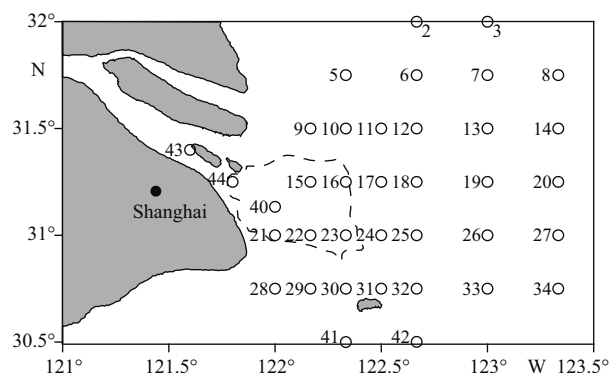
The Changjiang River estuary, which is about 120-km long and more than 90-km wide at its outer limit, has complex bottom topography. The TMZ is distributed widely across the estuary, and coincides locally with the mouth-bars. The pristine TMZ in the estuary was thought to be caused by sediment resuspension, which was primarily controlled by the interaction of tidal currents and river discharge (Li and Zhang, 1998; Pan et al., 1999). The TMZ was usually located at the edge of saline intrusion, and its magnitude and extent varies seasonally (Wu et al., 2012).

As one of the largest estuaries in the world, the partial pressure of CO<sub>2</sub> ( $p\text{CO}_2$ ) and CO<sub>2</sub> flux of the Changjiang River estuary has recently gained increased attention. Chen et al. (2008) measured the surface  $p\text{CO}_2$  in the environs of the estuary and the nearby Huangpu River outlet in the summer of 2003, using an underway-pumping system. However, their investigation was restricted to a measuring line. Zhai et al. (2007) examined the surface  $p\text{CO}_2$ , dissolved inorganic carbon and total alkalinity in the estuary, but their investigation focused on the inner and outer estuary, with only a few sampling stations in the middle. Zhai and Dai (2009) examined emphatically the surface  $p\text{CO}_2$  and dissolved oxygen (DO) in the outer Changjiang River estuary. Li et al. (2006b) measured the surface  $p\text{CO}_2$  in the estuary in May 2005, but only investigated five stations in the inner estuary. Gao et al. (2008) investigated  $p\text{CO}_2$  in the estuary, Hangzhou Bay and surrounding areas in August 2004, but almost all sampling stations were in Hangzhou Bay. At present, most investigations of the surface  $p\text{CO}_2$  in the estuary have not focused on the middle region of the estuary where the TMZ is situated. Thus, in this study, we focused on the air-sea CO<sub>2</sub> fluxes and influencing factors in the TMZ and surrounding area in the Changjiang River estuary.

## 2 MATERIAL AND METHOD

### 2.1 Sampling

Measures of TA, pH, and auxiliary data (salinity ( $S$ ), temperature ( $T$ ), wind speed, barometric pressure, etc.) were acquired in February, August, November 2010, and May 2012, representing winter, summer, fall and spring, respectively. Water column samples were collected with a Sea-Bird CTD rosette system equipped with 2.0-L Niskin bottles onboard the R/V



**Fig.1 The study area and site deployment**

The area surrounded by the broken line is the maximum turbidity zone from Shen and Pan (2001).

*Huchongyu 4936*. These samples covered almost the entire salinity gradient (Fig.1).

### 2.2 Analysis

The pH, TA,  $S$  and  $T$  of the samples were measured in situ. Salinity and  $T$  were measured using a combined Sea-Bird SBE 19plus CTD probe on the Rosette sampler. The pH was measured using Thermo Scientific Orion 3-Star Plus pH Meters with resolutions of 0.001 and relative accuracies of 0.002, and a Rex E-201-D combination electrode with measurement accuracy of 0.001. The electrode was calibrated at the beginning of every survey day by a three-point calibration using homemade buffers, including phthalate ( $\text{C}_8\text{H}_5\text{KO}_4$  0.05 mol/L, pH=4.002 at 25°C), phosphate ( $\text{KH}_2\text{PO}_4$  0.025 mol/L and  $\text{Na}_2\text{HPO}_4$  0.025 mol/L, pH=6.864 at 25°C), and borate ( $\text{Na}_2\text{B}_4\text{O}_7$  0.01 mol/L, pH=9.182 at 25°C). The precision of pH measurements was 0.002 pH units. Measurement of pH was carried out within minutes of sampling, to prevent CO<sub>2</sub> evasion and biological modifications of the sample.

Samples for total alkalinity were filtered through Whatman GF/F filters and analyzed onboard by Gran titration as quickly as possible, with a precision of approximately 0.1%–0.3% (Gao et al., 2008). Dickson standard reference materials were used for quality control in the total alkalinity measurements.

The  $p\text{CO}_2$  can be determined using either a direct or indirect method. The direct method first equilibrates the sample water with air in an equilibrator, and then the  $p\text{CO}_2$  is measured using infrared spectrometry or gas chromatography. This method has been used in many surveys and was thought to be the most accurate method to determine  $p\text{CO}_2$  in subsurface seawater. An underway  $p\text{CO}_2$  measurement system is the most

commonly used instrument for this measurement, and  $p\text{CO}_2$  measurements from the Changjiang River estuary have mostly been acquired from this instrument. However, the direct method does not work well in the maximum turbidity zone, because most equilibrators are not designed for working in turbid waters (blockage by suspended material). This is the main reason that there is little  $p\text{CO}_2$  data from TMZs. Therefore, an indirect method may be a better choice for measuring  $p\text{CO}_2$ . The indirect method calculates  $p\text{CO}_2$  from at least two of the following parameters: pH, TA, and total dissolved inorganic carbon (DIC). pH and TA are the most commonly used parameters. Many studies have proved that  $p\text{CO}_2$  can be accurately calculated from pH and TA (Frankignoulle and Borges, 2002; Gray et al., 2011). In this study,  $p\text{CO}_2$  was computed from measurements of pH and TA using the carbonic acid constant sets proposed by Cai and Wang (1998), the borate acidity constant from Millero (1979) and the  $\text{CO}_2$  solubility coefficient of Weiss (1974).

### 2.2.1 $\text{CO}_2$ fluxes ( $\text{FCO}_2$ )

The  $\text{CO}_2$  flux across the sea-air interface was estimated at each station using the equation:

$$F = k_s \Delta p\text{CO}_2,$$

where  $k$  is the gas transfer velocity (in  $\text{cm/h}$ ),  $s$  is the solubility of  $\text{CO}_2$  in seawater as a function of temperature and salinity taken from Weiss (1974), and  $\Delta p\text{CO}_2$  is the gradient in  $\text{CO}_2$  partial pressure between the seawater and the atmosphere, in  $\mu\text{atm}$ .

In estuaries, and particularly in macrotidal estuaries, the relationship between  $k$  and wind speed is usually site-specific and can be significantly affected by wind, water current, fetch, and turbidity (Abril et al., 2009). Therefore, we applied the most recent parameterization of Abril et al. (2009) to compute  $k$ . Wind speed has been recognized as the main force driving turbulence and gas exchange at the air-sea interface. As wind speed at one site varied greatly with time, the  $\text{CO}_2$  flux calculated by instantaneous wind speed was thus an instantaneous value, which may be an inaccurate representative of the region. The  $\text{CO}_2$  flux calculated by the average wind speed over a long period may have better results for this region. Therefore, in this paper, we used the average annual wind speed in different seasons to calculate the seasonal  $\text{CO}_2$  flux. The average wind speeds used were acquired from the pilot vessel station (122°04.12'E, 31°04.07'N) from 1979 to 1998

(Sui, 2003). Atmospheric  $p\text{CO}_2$  ( $p\text{CO}_{2\text{air}}$ ) was adopted from Zhai et al. (2007) and Zhai and Dai (2009). Tidal data for the periods of observations were collected at each station by a direct reading current meter (model SLC9-2).

Suspended particulate matter (SPM) concentration was determined as the weight of material retained on a Whatman GF/F membrane per volume unit after drying to constant weight at 60°C. The concentrations of chlorophyll *a* (Chl *a*) were determined spectrophotometry, after a 90% acetone extraction of the material retained on a Whatman GF/C membrane.

## 3 RESULT AND DISCUSSION

### 3.1 Spatial and temporal distributions of SPM

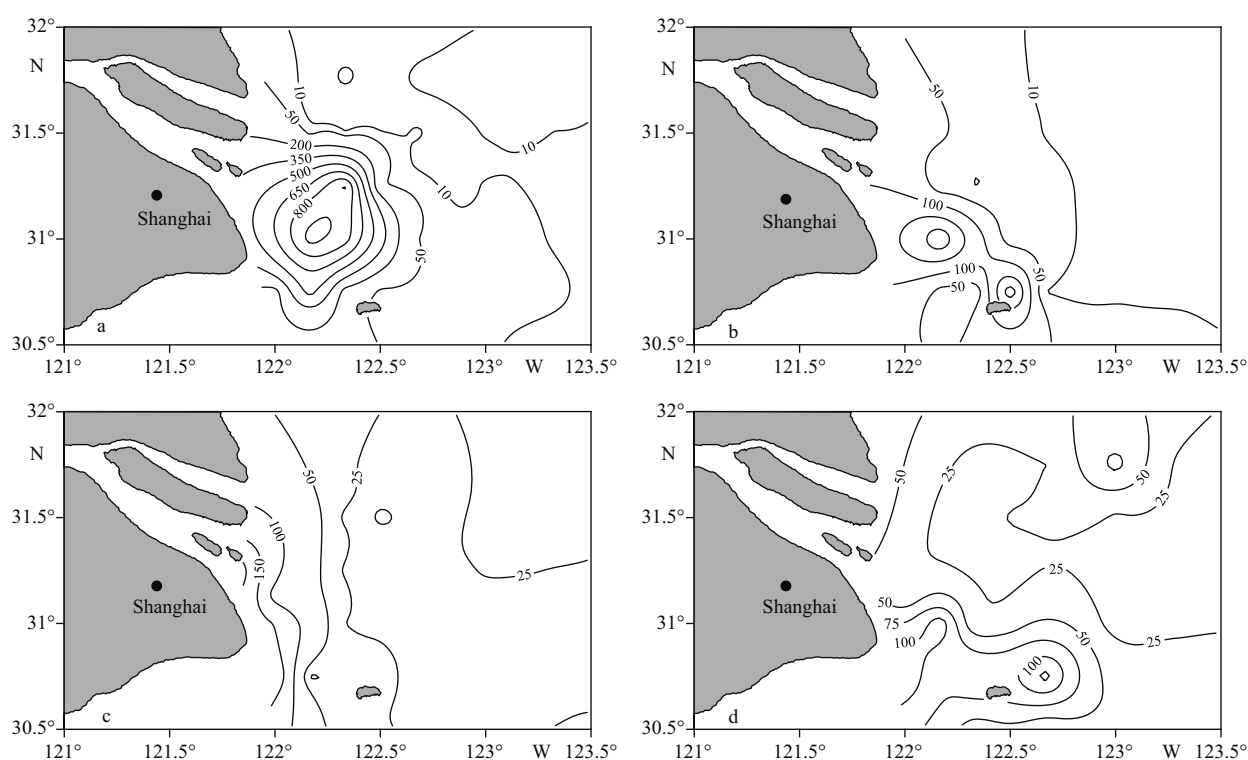
The SPM concentrations in the surface water of the TMZ and surrounding area in the Changjiang River estuary are shown in Fig.2 and Table 1. Similar to other investigations (Wu et al., 2012), the TMZ occurred in the south of the estuary. However, there were distinct seasonal differences in the magnitude and extent of the TMZ. The SPM concentration in the surface water of the TMZ and surrounding area was highest in February, with an average value of 177  $\text{mg/L}$ . This peak may be related to the strong winds at the time of sampling, which caused high levels of sediment to be re-suspended. It is known that the SPM in the bottom layer is commonly diffused to the upper layer during the dry season (Wu et al., 2012). The SPM concentration was lower in May and August, than in February and August. This pattern related to differences in transported matter from the Changjiang River and sea conditions between winter and summer.

### 3.2 Spatial and temporal distributions of TA

The overall TA distribution pattern in the surface water of the TMZ and surrounding area in the estuary did not vary seasonally, with TA increasing from the river mouth to the offshore region (Fig.3). However, there were small-scale differences in TA distribution patterns between seasons. The average and range of TA was similar in February and May (2 056  $\mu\text{mol/L}$ , Table 1). Minimum average TA was recorded in August (2 018  $\mu\text{mol/L}$ ), which was the Changjiang River flood season, with a large amount of fresh water entering the estuary and surrounding area. The TA was slightly higher in November than in August, but lower than in February and May.

**Table 1 Environmental parameters, CO<sub>2</sub> system and CO<sub>2</sub> fluxes in surface waters of the TMZ and surrounding area**

Item	May		August		November		February	
	Range	Average	Range	Average	Range	Average	Range	Average
TA (μmol/L)	1 646–2 315	2 056	1 681–2 147	2 018	1 729–2 241	2 034	1 780–2 381	2 056
pH	8.013–8.800	8.283	7.960–8.650	8.318	7.950–8.100	8.016	7.791–8.102	7.997
pCO <sub>2</sub> (μatm)	41–523	265	64–533	223	348–559	449	331–600	440
FCO <sub>2</sub> (mmol/(m <sup>2</sup> ·d))	-26–11	-9	-31–14	-16	-3–18	5	-5–20	5
Salinity	0.14–31.79	20.09	1.26–29.86	21.65	0.95–32.52	23.21	11.55–33.43	27.00
Water temperature (°C)	18.40–24.45	20.26	23.52–31.38	27.14	14.16–18.13	16.25	7.32–10.56	9.06
Wind speed (m/s)	-	5.47	-	5.54	-	5.02	-	5.22
Chl <i>a</i> (mg/m <sup>3</sup> )	0.43–2.7	1.38	0.58–2.75	1.36	0.70–1085	1.06	0.37–1.27	0.68
DO (mg/dm <sup>3</sup> )	7.48–16.34	9.34	5.61–8.43	7.21	-	-	8.98–10.30	9.63
Suspended matter (mg/dm <sup>3</sup> )	0.6–241	35	19–247	46	8.6–316	67	8.5–997	177

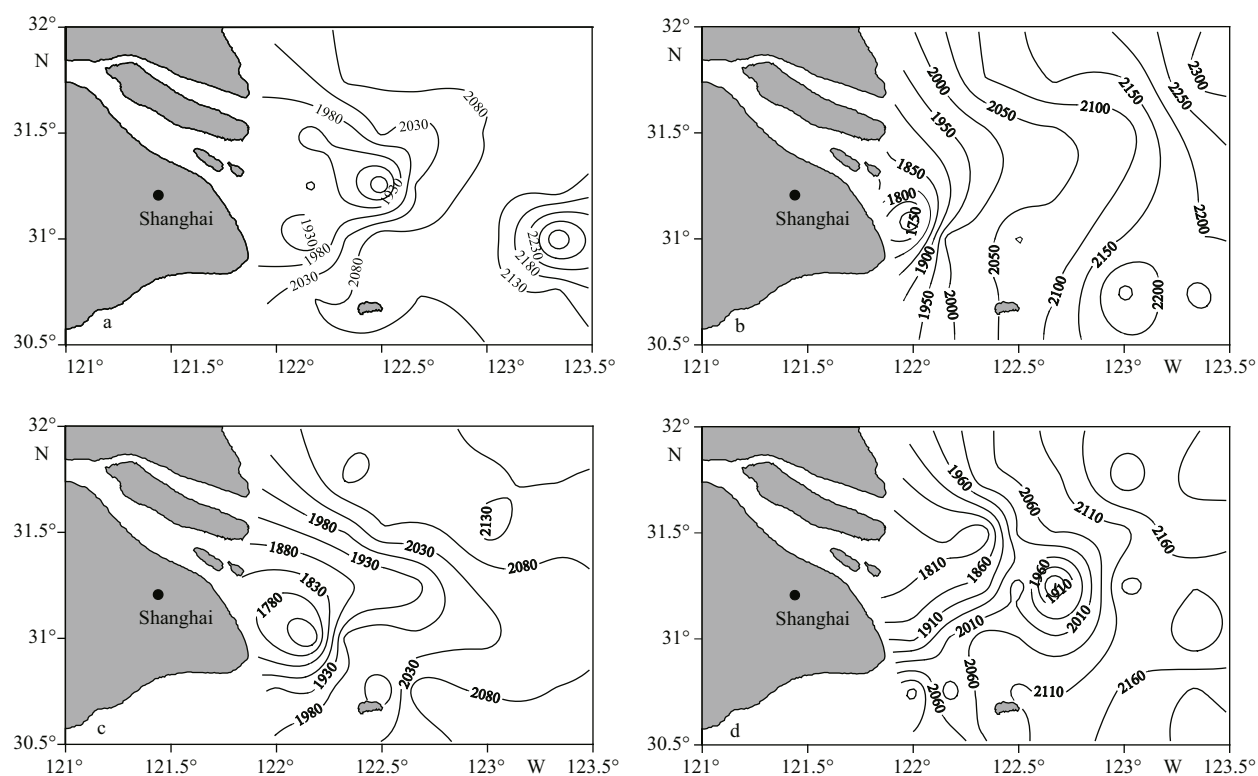
**Fig.2 Horizontal distributions of SPM concentrations in surface water of the TMZ and surrounding area in different seasons**

a. February; b. May; c. August; d. November.

### 3.3 Spatial and temporal distributions of pCO<sub>2</sub>

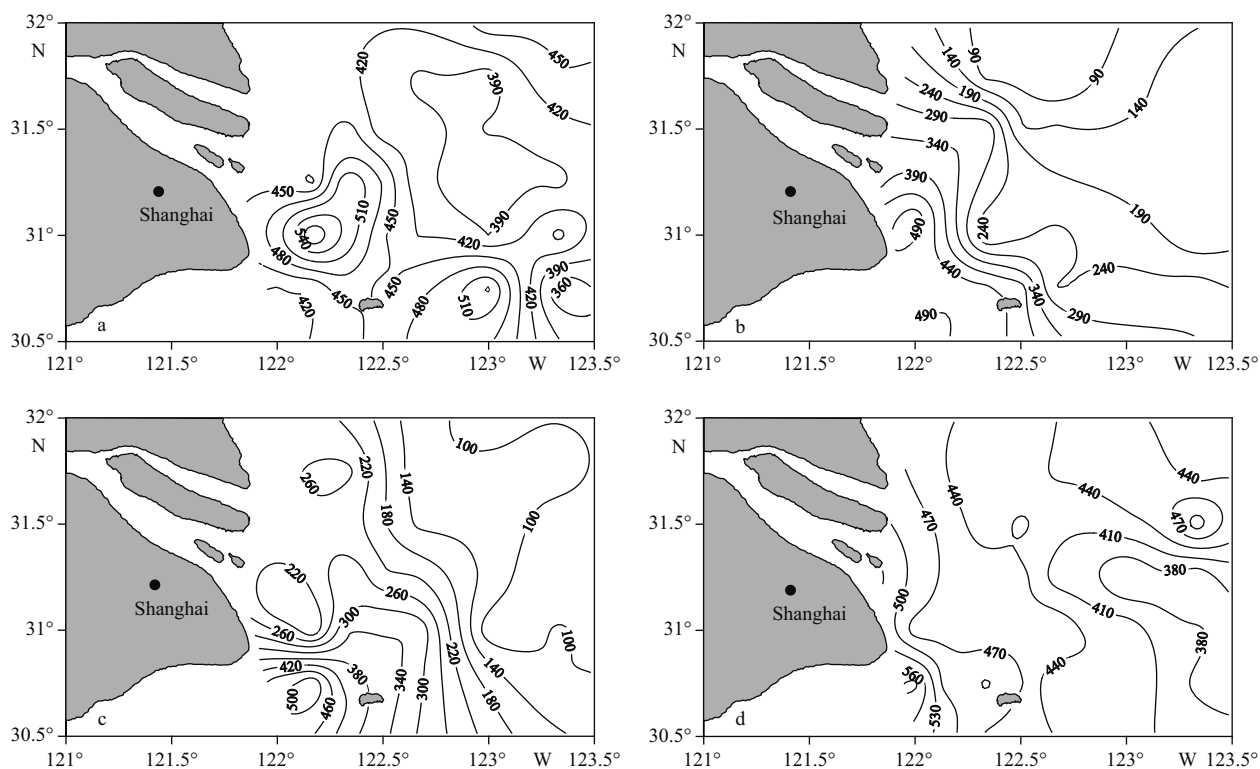
The pCO<sub>2</sub> distribution pattern in the surface water of the TMZ and surrounding area was more complicated than that of TA, with pCO<sub>2</sub> highly variable both spatially and seasonally during the four cruises (Fig.4). During our investigation periods, the largest pCO<sub>2</sub> (600 μatm) appeared in February and the lowest in May (40 μatm). In February and November, the surface pCO<sub>2</sub> was highest in the southwest of the region surveyed, decreased to its lowest in the center,

and then increased to the northwest. During fall and winter, surface pCO<sub>2</sub> was oversaturated with respect to the atmosphere in most regions. In February, pCO<sub>2</sub> ranged from 331 to 600 μatm with an average of 440 μatm in February. Similarly, in November, pCO<sub>2</sub> ranged from 348 to 559 μatm with an average of 449 μatm. In May and August, pCO<sub>2</sub> decreased from the southwest to the northeast of the region, and was unsaturated with respect to the atmosphere in most regions. In May, pCO<sub>2</sub> ranged from 41 to 523 μatm with an average of 265 μatm, with a range of 64–



**Fig.3** Horizontal distributions of TA in surface water of the TMZ and surrounding area in different seasons

a. February; b. May; c. August; d. November.

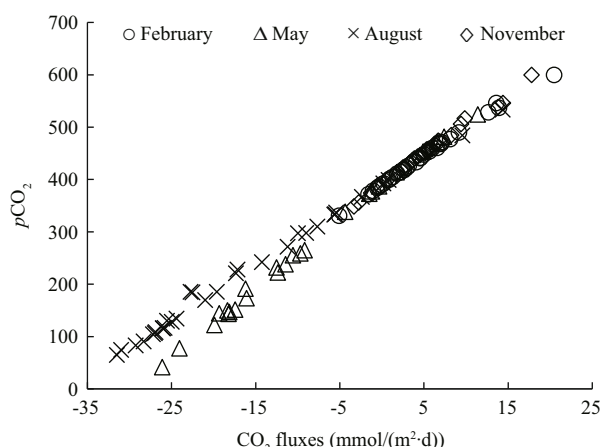


**Fig.4** Horizontal distributions of  $p\text{CO}_2$  in surface water of the TMZ and surrounding area in different seasons

a. February; b. May; c. August; d. November.



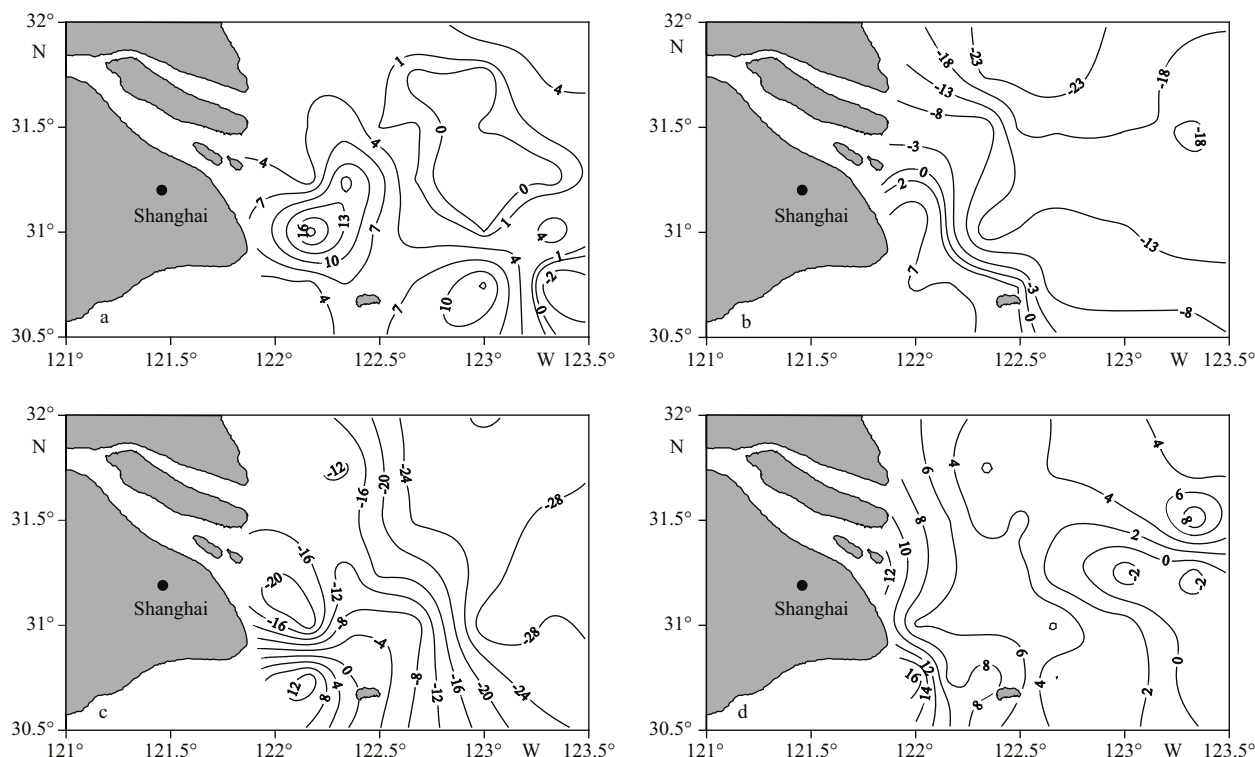
533  $\mu\text{atm}$  and an average of 223  $\mu\text{atm}$  in August. Thus, the surface  $p\text{CO}_2$  was higher in February and November, but lower in May and August, which is in contrast to the outer estuary, where there was much less seasonal variability (Zhai and Dai, 2009). Specifically, the outer estuary  $p\text{CO}_2$  winter range was 320–380  $\mu\text{atm}$  (average  $\sim 345$   $\mu\text{atm}$ ), spring range 180–450  $\mu\text{atm}$  (average  $\sim 330$   $\mu\text{atm}$ ), summer range 150–620  $\mu\text{atm}$  (average  $\sim 310$   $\mu\text{atm}$ ), and autumn range 120–540  $\mu\text{atm}$  (average  $\sim 375$   $\mu\text{atm}$ ).



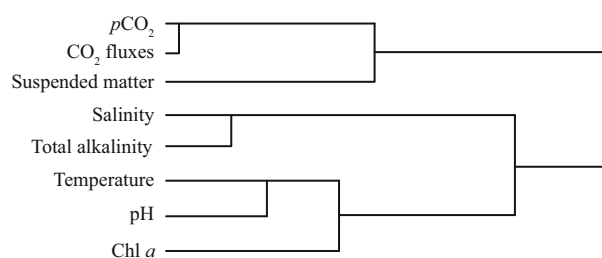
**Fig.5** Correlation of  $\text{FCO}_2$  and  $p\text{CO}_2$  in the TMZ and surrounding area

### 3.4 Spatial and temporal distributions of $\text{FCO}_2$

In the TMZ and surrounding area of the Changjiang River estuary,  $\text{FCO}_2$  and  $p\text{CO}_2$  were significantly correlated (Fig.5). Therefore,  $\text{FCO}_2$  distribution in this region was similar to that of  $p\text{CO}_2$ , showing clear seasonal variations (Fig.6). Overall, the TMZ and surrounding area acted as a source of atmosphere  $\text{CO}_2$  in February and November, but as a sink in May and August.  $\text{FCO}_2$  ranged between -5 and 20  $\text{mmol}/(\text{m}^2\cdot\text{d})$  in February and -3 and 18  $\text{mmol}/(\text{m}^2\cdot\text{d})$  in November. In the east central region,  $\text{FCO}_2$  was negative throughout the study period. In May and August,  $\text{FCO}_2$  was greatest in the northwestern region, then decreased in the northeastern region, and reached its lowest value in the north in May, or in the east in August. The sink or source characteristic of atmosphere  $\text{CO}_2$  in the TMZ was quite different to the outer estuary. First, in winter, the TMZ and surrounding area in the Changjiang River estuary was a source, but the outer estuary was a sink of atmosphere  $\text{CO}_2$  (Zhai and Dai, 2009). Second, in spring and summer, the TMZ and surrounding area was a stronger sink of atmosphere  $\text{CO}_2$  than the outer estuary. Third, the TMZ and surrounding area was a stronger source of atmosphere  $\text{CO}_2$  than the outer estuary in the autumn.



**Fig.6** Horizontal distributions of  $\text{FCO}_2$  across the seawater-air interface in the TMZ and surrounding area in different seasons  
a. February; b. May; c. August; d. November.



**Fig.7 Cluster analysis of the influencing factors in the TMZ and surrounding area of the Changjiang River estuary**

### 3.5 The influencing factors of FCO<sub>2</sub>

According to our study, the sink or source characteristics of atmosphere CO<sub>2</sub> were quite different in the TMZ and surrounding area, compared with the outer estuary. In an estuarine area, the TMZ acts as a filter between the terrestrial and marine realms, with most suspended particle carried by rivers deposited in the TMZ. Lisitsyn (1995) calculated that, worldwide, about 93%–95% of the suspended and 20%–40% of the dissolved riverine materials are deposited in what he calls the “marginal filter”. Shen et al. (2008) thought that the chemical behavior, transfer, and fate of organic matter might be affected by the TMZ. Therefore, it is not surprising that the characteristics of FCO<sub>2</sub> are different in the TMZ, compared with other regions. However, FCO<sub>2</sub> can be influenced by several factors, including the biological production or respiration of organic matter, calcification/dissolution of calcium carbonate, thermodynamic effects, physical mixing, and nitrification. However, the influences of these various factors are not isolated, rather they are interlinked. To identify the main influencing factors of CO<sub>2</sub> fluxes in the TMZ, we discuss seven of the said factors.

Cluster analysis of CO<sub>2</sub> fluxes and its influencing factors was used to determine which factor had the closest relationship with CO<sub>2</sub> flux (Fig.7). Cluster analysis classified the factors into three homogenous clusters, with each cluster influencing CO<sub>2</sub> fluxes in similar ways. Overall, *p*CO<sub>2</sub> concentration had the closest relationship with CO<sub>2</sub> fluxes, with a direct strong effect of *p*CO<sub>2</sub> on CO<sub>2</sub> flux, which accorded with the calculation of CO<sub>2</sub> fluxes from *p*CO<sub>2</sub> differences between seawater and air. Suspended matter concentration was also closely linked to CO<sub>2</sub> fluxes. The other influencing factors all had an indirect effect on CO<sub>2</sub> fluxes.

Factor analysis examined the influencing factors of CO<sub>2</sub> fluxes in the TMZ and surrounding area (Table 2). All of the influencing parameters can be represented

**Table 2 Factor analysis on the influencing factors of FCO<sub>2</sub> in the TMZ and surrounding area**

Factor	Eigenvalue	Percentage of variance	Percentage of accumulated variance
1	3.902	48.774	48.774
2	2.023	25.282	74.056
3	0.814	10.179	84.235
4	0.683	8.543	92.778
5	0.377	4.709	97.487
6	0.172	2.145	99.632
7	0.025	0.311	99.943
8	0.005	0.057	100.000

**Table 3 Factor loading matrix of factor analysis (after rotation)**

Parameter	1	2
Temperature	0.813	-0.213
Salinity	-0.061	0.950
Total alkalinity	0.067	0.895
<i>p</i> CO <sub>2</sub>	-0.946	-0.215
pH	0.950	0.101
Suspended matter	-0.396	-0.336
Chl <i>a</i>	0.586	-0.320
CO <sub>2</sub> flux	-0.949	-0.190

by two factors, which contribute 74% of all information. Factor 1 contributed far more than factor 2. Table 3 presents a factor-loading matrix for the factor analysis. Factor 1 mainly includes temperature, *p*CO<sub>2</sub>, and pH, suspended matter, and Chl *a*, with these factors having the greatest influence on CO<sub>2</sub> fluxes in the TMZ of the estuary. Factor 2 mainly includes salinity and alkalinity, which have a smaller impact on CO<sub>2</sub> fluxes. The importance of each influencing factor is discussed in the following section.

#### 3.5.1 Suspended matter in the TMZ

The TMZ is the region with higher suspended matter concentration than its surrounding area, and generally appears in estuaries. The location and strength of the TMZ usually changes with the estuarine circulation pattern. The TMZ, which is characterized by high suspended-matter concentrations, is an ideal site for physical, chemical, and biological reactions between dissolved and particulate species (Gebhardt et al., 2005). Therefore, the TMZ should play an important role in CO<sub>2</sub> degassing in estuaries.

**Table 4 Correlation between FCO<sub>2</sub> and its influencing factors**

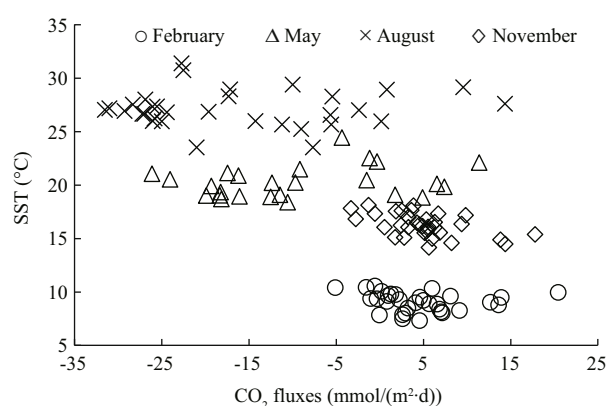
Season	SST <sup>a</sup>	Salinity	Alkalinity	pCO <sub>2</sub>	pH	Suspended particle	Chl <i>a</i>
February	0.150	-0.572**	-0.347	0.998**	-0.937**	0.621**	0.398*
May	0.251	-0.641**	-0.651**	0.999**	-0.931**	0.555**	-0.582**
August	0.031	-0.312	-0.280	0.998**	-0.969**	0.521**	0.052
November	-0.525**	-0.609**	-0.270	0.996**	-0.798**	0.504**	0.455**
Total year	-0.654**	-0.149	-0.203*	0.993**	-0.956**	0.349**	-0.403**

<sup>a</sup>: SST: sea surface temperature. \*: Correlation is significant at the 0.05 level (2-tailed). \*\*: Correlation is significant at the 0.01 level (2-tailed).

Suspended matter had a significant influence on the CO<sub>2</sub> fluxes in the TMZ and surrounding area, according to cluster analysis and the significant positive correlation between CO<sub>2</sub> fluxes and the concentration of suspended particles in this region (Table 4). The most conspicuous feature of the TMZ is its high concentration of suspended particles. The decomposition of organic carbon in suspended particles was the main reason that suspended particles had a significant influence on FCO<sub>2</sub> levels. For example, in the Seine estuary TMZ, about 20% of both POC and DOC were degraded by organisms (Garnier et al., 2008). In the Changjiang River estuary, organic carbon content in suspended particles was higher, with a range of 0.48%–0.69% (Li et al., 2006a). Organic carbon content was highest in the TMZ, decreasing upstream of the river mouth and lowest in the outer sea of the river mouth (Lin et al., 2009). Therefore, the large amount of organic carbon being decomposed in the TMZ of the Changjiang River estuary must affect the exchange of CO<sub>2</sub> between water and the atmosphere. The influence of organic carbon decomposition on FCO<sub>2</sub> was most significant in winter, when phytoplankton abundance was at its annual minimum, and thus the influence of phytoplankton on FCO<sub>2</sub> was limited. This low phytoplankton abundance might be the reason that the TMZ acted as a source in winter, but the outer estuary acted as a sink.

### 3.5.2 Sea surface temperature (SST)

Temperature is one of the most important thermodynamic factors controlling the fluxes of CO<sub>2</sub>. Many researchers have reported that sea surface temperature (SST) has a positive correlation with seawater pCO<sub>2</sub> and CO<sub>2</sub> fluxes. This is because SST affects the equilibrium constants of dissolved inorganic carbon and, in particular, the solubility coefficient of CO<sub>2</sub>. Solubility of CO<sub>2</sub> in seawater decreases with an increase in SST, so pCO<sub>2</sub> rises with an increase in SST. It is reported that pCO<sub>2</sub> may rise by ~4% when temperature increases by 1°C (Borge

**Fig.8 Relationship between SST and FCO<sub>2</sub> in the TMZ and surrounding area**

and Frankignoulle, 2002). However, in the TMZ and surrounding area, SST had a significant negative correlation with CO<sub>2</sub> fluxes on intra-annual timescales (Table 4; Fig.8), suggesting that other factor(s) might overwhelm temperature in controlling the annual fluxes of CO<sub>2</sub>. Similar negative relationships between temperature and pCO<sub>2</sub> have been reported by Zhang et al. (2010) in the southern Yellow Sea in March, and by Chou et al. (2011) in the East China Sea during winter. Zhang et al. (2010) suggested that this correlation strongly indicated that the seasonal variations in pCO<sub>2</sub> were controlled by biological activity. There were no significant correlations between SST and FCO<sub>2</sub> during spring, summer, and autumn in the TMZ of the estuary in our study (Table 1; Fig.8). The results concur with Zhai and Dai's (2009) investigations in the outer estuary, where the relationship between SST and pCO<sub>2</sub> was generally random. In the outer estuary, SST changed very little but CO<sub>2</sub> flux was greatly variable during the study period, which indicated that temperature had not significantly influenced FCO<sub>2</sub>. However, SST had a significant positive correlation with Chl *a* on intra-annual timescales (Table 1; Fig.9). The SST could have influenced the FCO<sub>2</sub> by controlling phytoplankton growth; phytoplankton biomass may increase as SST



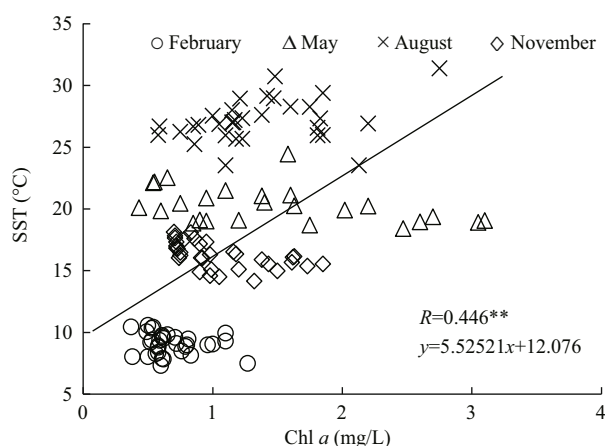


Fig.9 Relationship between SST and Chl *a* in the TMZ and surrounding area

increases, and thus use more  $\text{CO}_2$  in seawater. The  $p\text{CO}_2$  in seawater may thus decrease, leading to a negative relationship between SST and  $\text{FCO}_2$ .

### 3.5.3 Phytoplankton

In theory, phytoplankton can absorb  $\text{CO}_2$  and release  $\text{O}_2$ , and thus phytoplankton blooms will decrease the  $\text{CO}_2$  concentration in seawater. Therefore, the relationship between phytoplankton and  $p\text{CO}_2$  in seawater should be negative. In the TMZ and surrounding area,  $\text{FCO}_2$  had a significant negative relationship with Chl *a* throughout the year (Table 4). As shown in Table 1, in the TMZ and surrounding area, Chl *a* concentration was high in May and August and low in November and February. The results are consistent with those of Song et al. (2009) and Wu et al. (2004). Song et al. (2009) showed that Chl *a* concentration in the estuary changed seasonally, with higher levels in spring and summer and lower levels in autumn and winter. Wu et al. (2004) showed that the peak period of phytoplankton abundance coincided with that of Chl *a*, which occurred during summer, but the abundance of phytoplankton was very low during the winter (dry season). Therefore, the TMZ and surrounding area acted as a sink of atmospheric  $\text{CO}_2$  in summer and spring, but as a source in winter and autumn. However, there was no clear negative correlation between phytoplankton abundance and  $\text{FCO}_2$  in every month, bar August (Table 4), which may be linked to phytoplankton distribution in the TMZ of the estuary and surrounding area.

The phytoplankton distribution in the TMZ and surrounding area was obviously influenced by the tides. The abundance of phytoplankton was larger during the spring tide than the neap tide, in both the

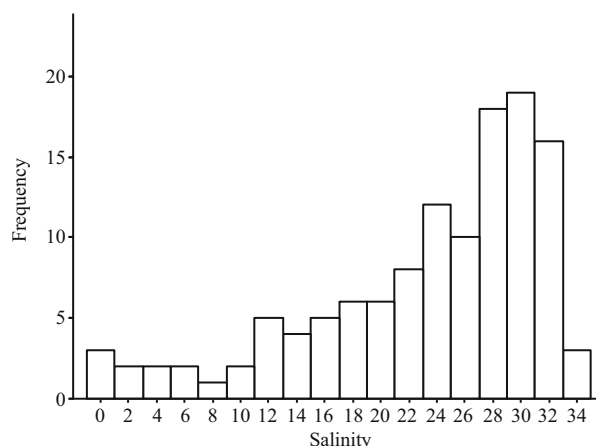
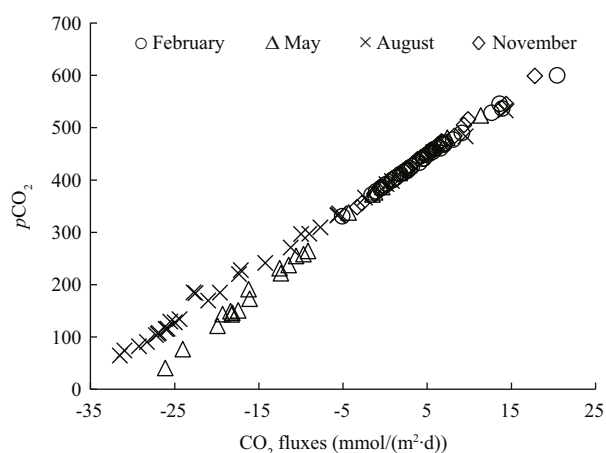


Fig.10 Surface water salinity frequencies in the TMZ and surrounding area

flood and dry seasons. The abundance of phytoplankton was larger in the tidal flood than the tidal ebb, both during spring and neap tides (Gu et al., 1995). Additionally, phytoplankton are known to have patchy distributions in the TMZ of estuaries (Gu et al., 1995). The variable correlation between Chl *a* and  $\text{FCO}_2$  may be caused by the asymmetric distribution of phytoplankton and the inconsistent tide time during sampling.

### 3.5.4 Salinity

In estuaries, salinity is one of the important factors controlling the flux of  $\text{CO}_2$ , especially in the mixed zone, where  $\text{FCO}_2$  decreases with increasing salinity. The TMZ of the estuary was located in the freshwater-seawater mixing zone. Further, salinity was less than 32 in most sampled sites throughout the four sampling seasons (Fig.10). Thus, the salinity in the TMZ and surrounding area should play an important role in the exchange of  $\text{CO}_2$  across the sea-air interface. Table 4 shows the relationship between sea surface salinity (SSS) and the flux of  $\text{CO}_2$ . There is a clear negative correlation between  $\text{FCO}_2$  and SSS in February, May and November, suggesting that low  $\text{FCO}_2$  values were mainly influenced by seasonal variations in SSS. No such correlation was observed in August. In February, May, and November, mixing between the low-temperature/high- $p\text{CO}_2$  estuarine or coastal water and high-temperature/low- $p\text{CO}_2$  water of the East China Sea influenced the  $\text{FCO}_2$ . However, the weak relationship between  $\text{FCO}_2$  and salinity revealed that salinity was not the major influencing factor on  $\text{FCO}_2$  in summer. According to the seasonal variation in primary production, low  $\text{FCO}_2$  in summer was controlled by high primary production.



**Fig.11 The relationship between CO<sub>2</sub> fluxes and pH in the TMZ in surface water of the TMZ and surrounding area**

### 3.5.5 Total alkalinity and pH

The weak negative relationship between FCO<sub>2</sub> and total alkalinity indicated that total alkalinity was not the major controlling factor of FCO<sub>2</sub> in the TMZ and surrounding area (Fig.11 and Table 4). However, the pH in the surface water had a significant negative correlation with FCO<sub>2</sub>, which indicated pH had a significance influence on FCO<sub>2</sub>. The main alkaline compounds in the water are bicarbonate, carbonate, and hydroxide, which remove H<sup>+</sup> ions, and more than 90% of them are bicarbonates (Song, 2009). The pH is measured on a logarithmic scale and measures the amount of hydrogen ions (H<sup>+</sup>) in the water; a 0.1 reduction means that the quantity of H<sup>+</sup> ions has increased by 30%. The carbonate system exists in the following equilibrium in water:



When the [H<sup>+</sup>] or [HCO<sub>3</sub><sup>-</sup>] increase, the equilibrium may move to the left, the pCO<sub>2</sub> in the seawater will increase, and the uptake capacity of atmospheric CO<sub>2</sub> will reduce. As bicarbonate concentration is higher in seawater than the atmosphere, the pCO<sub>2</sub> variation caused by its change is not significant, and thus total alkalinity has a weak correlation with FCO<sub>2</sub>. However, any little increase (or decrease) of pH may lead to a notable decrease (or increase) in [H<sup>+</sup>] and cause a similar variation in pCO<sub>2</sub> with it. Therefore, pH has a strong negative correlation with FCO<sub>2</sub>. In the TMZ and surrounding area, total alkalinity was mainly between 2 000 and 2 200 μmol/L throughout the four seasons, so any changes in total alkalinity were relatively small. Thus, although total alkalinity had a weak influence on FCO<sub>2</sub>, the large change in pH from

7.791 to 8.800, indicates a large change in [H<sup>+</sup>], so pH had a significant negative correlation with FCO<sub>2</sub>.

## 4 CONCLUSION

The turbidity maximum zone (TMZ) in estuaries can differ in terms of its role as a source or sink of atmospheric CO<sub>2</sub> because of its high concentration of suspended matter. The TMZ and surrounding area in the Changjiang River estuary acted as a source of atmosphere CO<sub>2</sub> in February and November, and as a sink in May and August. The average FCO<sub>2</sub> was -9, -16, 5, and 5 mmol/(m<sup>2</sup>·d) in May, August, November, and February, respectively. However, the source/sink characteristic differed in the TMZ compared with the outer estuary, especially in the winter, when the TMZ and surrounding area was a source, but the outer estuary was a sink of atmospheric CO<sub>2</sub>. In addition to suspended matter, phytoplankton production and pH were the main factors influencing FCO<sub>2</sub> in the TMZ and surrounding area. The influence of temperature, salinity and total alkalinity on the FCO<sub>2</sub> was weak in the TMZ and surrounding area. Organic carbon decomposition in suspended matter was the main reason the TMZ and surrounding area was a CO<sub>2</sub> source in winter, and phytoplankton production was the main reason it was a CO<sub>2</sub> sink in summer.

## References

- Abril G, Commarieu M V, Sottolichio A, Bretel P, Guerin F. 2009. Turbidity limits gas exchange in a large macrotidal estuary. *Estuarine Coastal and Shelf Science*, **83**(3): 342-348.
- Borges A V, Dellile B, Frankignoulle M. 2005. Budgeting sinks and sources of CO<sub>2</sub> in the coastal ocean: diversity of ecosystems counts. *Geophysical Research Letters*, **32**: L14601.
- Borges A V. 2005. Do we have enough pieces of the jigsaw to integrate CO<sub>2</sub> fluxes in the coastal ocean? *Estuaries*, **28**: 3-27.
- Cai W J, Dai M H, Wang Y C. 2006. Air-sea exchange of carbon dioxide in ocean margins: a province-based synthesis. *Geophysical Research Letters*, **33**(12): L12603.
- Cai W J, Wang Y C. 1998. The chemistry, fluxes, and sources of carbon dioxide in the estuarine waters of the Satilla and Altamaha Rivers, Georgia. *Limnology Oceanography*, **43**: 657-668.
- Cai W J. 2011. Estuarine and coastal ocean carbon paradox: CO<sub>2</sub> sinks or sites of terrestrial carbon incineration? *Annual Review of Marine Science*, **3**: 123-145.
- Chen C T A, Borges A V. 2009. Reconciling opposing views on carbon cycling in the coastal ocean: continental shelves as sinks and near-shore ecosystems as sources of atmospheric CO<sub>2</sub>. *Deep-Sea Research II*, **56**: 578-590.

- Chen C T A, Huang T H, Chen Y C, Bai Y, He X, Kang Y. 2013. Air-sea exchanges of CO<sub>2</sub> in the world's coastal seas. *Biogeosciences*, 6 509-6 544.
- Chen C T A, Huang T H, Fu Y H, Bai Y, He X Q. 2012. Strong sources of CO<sub>2</sub> in upper estuaries become sinks of CO<sub>2</sub> in large river plumes. *Current Opinion in Environmental Sustainability*, 4(2): 179-185.
- Chen C T A, Zhai W D, Dai M H. 2008. Riverine input and air-sea CO<sub>2</sub> exchanges near the Changjiang (Yangtze River) estuary: status quo and implication on possible future changes in metabolic status. *Continental Shelf Research*, 28: 1 476-1 482.
- Chou W C, Gong G C, Tseng C M, Sheu D D, Hung C C, Chang L P, Wang L W. 2011. The carbonate system in the East China Sea in winter. *Marine Chemistry*, 123: 44-55.
- Frankignoulle M, Borges A V. 2002. Direct and indirect pCO<sub>2</sub> measurements in a wide range of pCO<sub>2</sub> and salinity values (The Scheldt estuary). *Aquatic Geochemistry*, 7: 267-273.
- Gao X L, Song J M, Li X G, Li N, Yuan H M. 2008. pCO<sub>2</sub> and carbon fluxes across sea-air interface in the Changjiang estuary and Hangzhou Bay. *Chinese Journal of Oceanology and Limnology*, 26(3): 289-295.
- Garnier J, Billen G, Even S, Etcheber H, Servais P. 2008. Organic matter dynamics and budgets in the turbidity maximum zone of the Seine estuary (France). *Estuarine, Coastal and Shelf Science*, 77: 150-162.
- Gebhardt A C, Schoster F, Gaye-Haake B, Beeskow B, Rachold V, Unger D, Ittekkot V. 2005. The turbidity maximum zone of the Yenisei River (Siberia) and its impact on organic and inorganic proxies. *Estuarine, Coastal and Shelf Science*, 65: 61-73.
- Gray S E C, DeGrandpre M D, Moore T S, Martz T R, Friederich G E, Johnson K S. 2011. Applications of in situ pH measurements for inorganic carbon calculations. *Marine Chemistry*, 125: 82-90.
- Gu X G, Yuan Qi, Shen H T, Zhou Y Q. 1995. The ecological study on phytoplankton in maximum turbid zone of Changjiang estuary. *Journal of Fishery Sciences of China*, 2(1): 16-27. (in Chinese with English abstract)
- Guo X H, Dai M H, Zhai W D, Cai W J, Chen B S. 2009. CO<sub>2</sub> flux and seasonal variability in a large subtropical estuarine system, the Pearl River estuary, China. *Journal of Geophysical Research*, 114: G03013.
- Jiang L Q, Cai W J, Wang Y C. 2008. A comparative study of carbon dioxide degassing in river- and marine-dominated estuaries. *Limnology Oceanography*, 53(6): 2 603-2 615.
- Li J F, Zhang C. 1998. Sediment resuspension and implications for turbidity maximum in the Changjiang estuary. *Marine Geology*, 148: 117-124.
- Li X G, Song J M, Niu L F, Yuan H M, Li N, Gao X L. 2007. Role of the Jiaozhou Bay as a source/sink of CO<sub>2</sub> over a seasonal cycle. *Scientia Marina*, 71(3): 441-450.
- Li X G, Song J M, Yuan H M. 2006a. Inorganic carbon of sediments in the Yangtze River estuary and Jiaozhou Bay. *Biogeochemistry*, 77: 177-197.
- Li Y, Zhang L J, Su Z, Wang X L, Pan J M. 2006b. Rapid variation of pCO<sub>2</sub> and its influencing factors at the initial mixing stage of freshwater and saltwater in the Changjiang River estuary. *Periodical of Ocean University of China*, 36(2): 295-298. (in Chinese with English abstract)
- Lin J, Wu Y, Zhang J, Zhu Z Y. 2009. The primary study on the influence of the turbidity maximum zone on the organic carbon in the Changjiang. *Strait Science*, 30(6): 150-158. (in Chinese)
- Lisitsyn A P. 1995. The marginal filter of the ocean. *Oceanology*, 34: 671-682.
- Millero F J. 1979. The thermodynamics of the carbonate system in seawater. *Geochemica et Cosmochemica Acta*, 43: 1 651-1 661.
- Pan D A, Shen H T, Mao Z C. 1999. Formation mechanism and features of the turbidity maximum in the Changjiang River estuary. *Acta Oceanologica Sinica*, 21(4): 62-69. (in Chinese with English abstract)
- Shen H T, Pan D A. 2001. Turbidity Maximum in the Changjiang Estuary. China Ocean Press, Beijing, China. p.39. (in Chinese)
- Shen Z, Zhou S, Pei S. 2008. Transfer and transport of phosphorus and silica in the turbidity maximum zone of the Changjiang estuary. *Estuarine, Coastal and Shelf Science*, 78: 481-492.
- Song J M. 2009. Biogeochemical Processes of Biogenic Elements in China Marginal Seas. Springer-Verlag GmbH & Zhejiang University Press, Hangzhou, China. 662p.
- Song S Q, Sun J, Yu Z M. 2009. Vertical pattern of chlorophyll *a* in the Yangtze River estuary and its adjacent waters. *Chinese Journal of Plant Ecology*, 33(2): 369-379. (in Chinese with English abstract)
- Sui H B. 2003. Statistical Characteristics of Wave Distribution and Double Peak Spectrum in the Yangtze River Estuary. Dissertation for Master Degree Ocean University of China. (in Chinese)
- Weiss R F. 1974. Carbon dioxide in water and seawater: the solubility of a non-ideal gas. *Marine Chemistry*, 2: 203-215.
- Wu J X, Liu J T, Wang X. 2012. Sediment trapping of turbidity maxima in the Changjiang estuary. *Marine Geology*, 303-306: 14-25.
- Wu Y L, Fu Y N N, Zhang Y S, Pu X M, Zhou C X. 2004. Phytoplankton distribution and its relation to the runoff in the Changjiang (Yangtze) estuary. *Oceanologia et Limnologia Sinica*, 35(3): 246-251. (in Chinese with English abstract)
- Zhai W D, Dai M H, Guo X H. 2007. Carbonate system and CO<sub>2</sub> degassing fluxes in the inner estuary of Changjiang (Yangtze) River, China. *Marine Chemistry*, 107: 342-356.
- Zhai W D, Dai M H. 2009. On the seasonal variation of air-sea CO<sub>2</sub> fluxes in the outer Changjiang (Yangtze River) estuary, East China Sea. *Marine Chemistry*, 117: 2-10.
- Zhang L J, Xue L, Song M Q, Jiang C B. 2010. Distribution of the surface partial pressure of CO<sub>2</sub> in the southern Yellow Sea and its controls. *Continental Shelf Research*, 30: 293-304.