

Removal of phenol by *Isochrysis galbana* in seawater under varying temperature and light intensity*

LI Hao¹, MENG Fanping^{1, 2, **}, WANG Yuejie¹, LIN Yufei³

¹ Key Laboratory of Marine Environment and Ecology, Ministry of Education, Qingdao 266100, China

² College of Environmental Science and Engineering, Ocean University of China, Qingdao 266100, China

³ National Marine Hazard Mitigation Service, Ministry of Natural Resources of the People's Republic of China, Beijing 100194, China

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Abstract Phenol is a common industrial chemical produced and transported worldwide largely. Therefore, accidental spillage of phenol in the ocean causes an increasing concern. Microalgae are promising to remove phenol from marine waters. However, temperature and light intensity are two main factors that markedly influence biodegradation in marine environments. In this study, a marine golden alga *Isochrysis galbana* is selected to research the removal of phenol under different temperatures (10–30°C) and light intensities (0–240 $\mu\text{mol}/(\text{m}^2\cdot\text{s})$). The results show that the most suitable temperature and light intensity for phenol removal are 20°C and 180 $\mu\text{mol}/(\text{m}^2\cdot\text{s})$, respectively, and 100 mg/L of phenol can be completely removed by microalga in 24 h at these conditions. *I. galbana* can also remove phenol under dark and low-temperature conditions. The removal of phenol by *I. galbana* at diverse temperatures and light intensities conform to first-order kinetics, and the process under dark conditions conform to zero-order kinetics. Thus, *I. galbana* can be used in the in-situ bioremediation of polluted seawater by phenol.

Keyword: phenol; *Isochrysis galbana*; temperature; light intensity; biodegradation.

1 INTRODUCTION

Phenol is an important chemical raw material, and it is widely used in chemical, pharmaceutical, petroleum, leather, and other industries. It has an annual output of approximately 7 million tons (Senthilvelan et al., 2014). Phenol is one of the most common hazardous noxious substances (HNS) involved in accidental marine spills due to its high market demand and transportation volume (Cunha et al., 2015; Duan et al., 2017). For example, a total of 400 tonnes of phenol leaked into the seawater in the Port of Gothenburg, Sweden in 1973, because a cistern into which the phenol was loaded was suddenly ruptured (HELCOM, 2002). Phenol is soluble in water (Massalha et al., 2010), and it is highly toxic to aquatic organisms (Calabrese and Kenyon, 1991). Phenol is currently listed as a priority pollutant by the US Environmental Protection Agency (USEPA) (Du et al., 2009). French McCay et al. (2006) used prediction models to show that phenol is one of the

most dangerous chemicals for aquatic organisms used in bulk transport.

Conventional physical and chemical environmental restoration methods are not effective after the leakage of dissolved HNS in large-area waters, such as an ocean. The only thing that can be done for an HNS leakage is to take measures to speed up the natural dilution process. Microalgae are a primary producer in marine ecological environments, and they are an important part of the marine ecosystem. Microalgae can not only maintain autotrophic growth using photosynthesis but also utilize organic matter from the environment as its carbon source for heterotrophic growth (Perez-Garcia et al., 2011). Many marine microalgae can biodegrade specific pollutants,

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** Corresponding author: mengfanping@ouc.edu.cn

especially for phenols (Yang et al., 2002; Lima et al., 2004; Gao et al., 2011; Gao and Chi, 2015; Wang et al., 2019). Thus, microalgae can be used to remove organic contaminants from seawater due to its significant advantages, including being driven by solar energy, environmental friendliness, its role in the fixation and turnover of carbon, cheap food for animals (Wijffels et al., 2013; Matamoros et al., 2015; Xiong et al., 2017). However, the environmental conditions of oceans are complicated and diverse, and changing marine environmental conditions have a great impact on the removal of organic pollutants by microorganisms (Caracciolo et al., 2015). Temperature and light intensity are two basic factors that will affect biodegradation by microorganisms to a large extent (Bradley and Writer, 2014; Zhou et al., 2015; Surkatti and El-Naas, 2018).

Isochrysis galbana (a common feed microalga used widely in aquaculture) has been screened by Wang et al. (2019) as an excellent phenol-degrader. In this study, the removal of phenol by *I. galbana* is studied under different temperatures and light intensities in seawater. The kinetics is also studied to discuss their influence on the removal of phenol by temperature and light.

2 MATERIAL AND METHOD

2.1 Chemicals, microalga source, and culture medium

Phenol was obtained from the SinoPharm Chemical Reagent Co., Ltd. (Shanghai, China). All of the other solvents and reagents were analytical grade. Natural seawater was used in this study, which was taken from the coastal waters of Qingdao, China. Seawater was filtered using a 0.45- μm microporous membrane, and then sterilized under 120°C for 20 min.

Golden alga (*Isochrysis galbana* MACC/H59) was obtained from the algal culture collection at the Ocean University of China (OUC). The pure-cultured microalga was cultivated in a sterilized F/2 medium (Guillard, 1975) with the following composition (L): NaNO_3 , 75 mg; $\text{Na}_2\text{HPO}_4 \cdot \text{H}_2\text{O}$, 5 mg; $\text{EDTA} \cdot \text{Na}_2$, 4.36 mg; a solution of trace elements ($\text{ZnSO}_4 \cdot 4\text{H}_2\text{O}$, 0.023 g/L; $\text{CoCl}_2 \cdot 6\text{H}_2\text{O}$, 0.012 g/L; $\text{FeCl}_3 \cdot 6\text{H}_2\text{O}$, 3.2 g/L; $\text{MnCl}_2 \cdot 4\text{H}_2\text{O}$, 0.178 g/L; $\text{CuSO}_4 \cdot 5\text{H}_2\text{O}$, 0.010 g/L; $\text{Na}_2\text{MoO}_4 \cdot 2\text{H}_2\text{O}$, 0.006 g/L), 1 mL; and a solution of vitamins (Vitamin B12, 0.000 5 g/L; Vitamin B1, 0.100 g/L; biotin, 0.000 5 g/L) 1 mL. The experiments and sampling processes were carried out in a superclean bench, and the flasks were sealed

with ventilated membranes (pore-size is 0.45 μm) to ensure air entry and axenic environments. The microalga was cultivated to the exponential phase for next-step inoculation operation.

2.2 The effect of temperature on the growth of microalga and the biodegradation of phenol

The experiment was conducted in light incubators. The initial inoculation density of each microalga cell was 5×10^5 cells/mL. Each 100-mL culture was cultivated in a 250-mL flask. There were five temperature groups, and each group contained 0 (control), 50, and 100 mg/L of phenol (50 mg/L of phenol is higher than LC_{50} values of most marine fishes and crustaceans (Duan et al., 2018), so phenol above 50 mg/L can be considered as a serious hazard to the marine environment). There are three repetitions in each temperature. First, all of the groups were pre-incubated in an incubator for 24 h at a temperature of 20°C, 60 $\mu\text{mol}/(\text{m}^2 \cdot \text{s})$ of light intensity, and a 14-h light/10-h dark cycle. Then the five groups were incubated at temperatures of 10°C, 15°C, 20°C, 25°C, and 30°C. The other factors at this stage were the same as during pre-incubation. The flasks were shaken with a speed of 150 r/min. The density of the microalga and concentration of phenol were measured every 2 h.

2.3 The effect of light intensity on the growth of microalga and biodegradation of phenol

There were five groups established to investigate the effects of light intensity. Each group contained 0 (control), 50, and 100 mg/L of phenol. The inoculum and pre-incubation was the same as before. Afterward, the five groups were incubated in 0 (dark), 60, 120, 180, and 240 $\mu\text{mol}/(\text{m}^2 \cdot \text{s})$ of light intensity, and there are three repetitions in each illumination. In addition, one group without inoculum (containing 50 and 100 mg/L of phenol) was set as a control of photolysis in each illumination. Other factors were the same as during pre-incubation. The density of the microalga and the concentration of phenol were measured every 2 h.

2.4 The effect of kinetics on the biodegradation of phenol by *I. galbana*

The exponential rate model is widely applied to the removal process of organics as a result of its good applicability and fitting effect (Liu et al., 2014). There are zero-order model, first-order model and

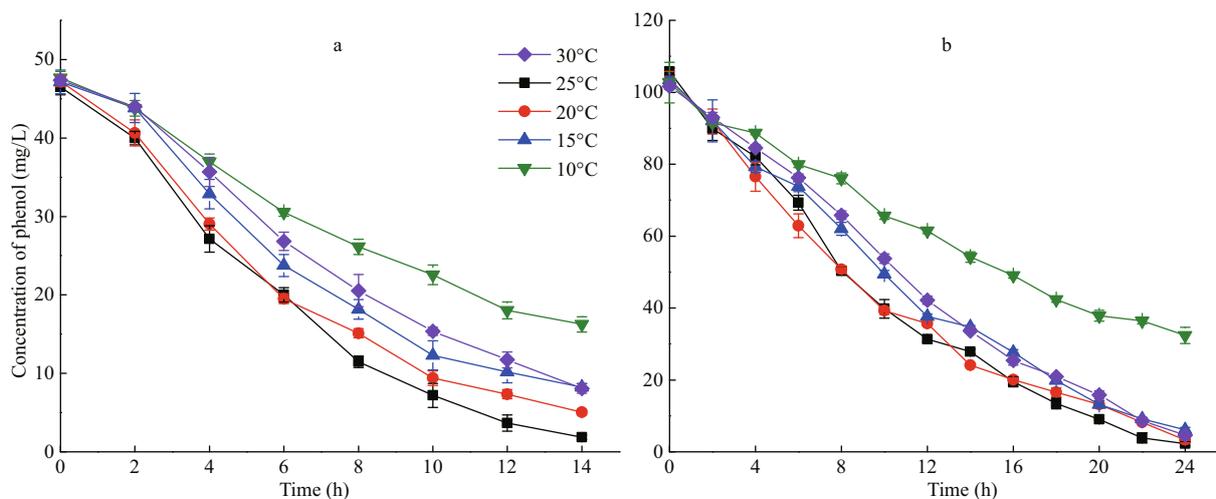


Fig.1 Removal of 50 mg/L (a) and 100 mg/L (b) of phenol using *I. galbana* under different temperatures

pseudo first-order model widely used for modeling the degradation of organic contaminants (Saravanan et al., 2009; Wan et al., 2010; Silambarasan and Abraham, 2013). The zero-order model and pseudo first-order model were used to describe the degradation kinetics of phenol removal in this study (Saravanan et al., 2009; Wang et al., 2015), as the following Eqs.1 & 2:

$$S_t = S_0 - Kt, \quad (1)$$

$$S_t = S_0 - S_0 \times e^{-Kt}, \quad (2)$$

where S_0 is the initial concentration of phenol (mg/L); and S_t is the concentration of phenol (mg/L) at time t . In order to estimate kinetic parameters for phenol removal, the Eqs.1 & 2 were solved using the linear and nonlinear regression analysis, respectively. The attenuation concentration of phenol was determined by Eq.3:

$$S_a = S_0 - S_t, \quad (3)$$

where S_a is the attenuated concentration of phenol (mg/L). In zero-order kinetics, t is the half-life when $S_t/S_0 = 1/2$. In first-order kinetics, the half-life ($t_{1/2}$) of phenol removal can be measured using Eq.4 (Sinkkonen and Paasivirta, 2000):

$$t_{1/2} = \ln 2 / K. \quad (4)$$

2.5 Determination of the cell density of microalgae and phenol concentration

Cell densities of microalga were measured using a haemocytometer.

The concentrations of phenol were analyzed using the 4-aminoantipyrine spectrophotometric method at 510 nm (Emerson, 1943). Before measuring the phenol concentrations, a 2-mL sample was withdrawn

from each flask and centrifuged at 4 400×g for 5 min and then filtered through a 0.45- μ m filter. The supernatants were used to monitor the residual phenol concentrations.

2.6 Statistical analysis

The means and standard deviations of all the results were obtained from three independent replicates. Statistical analysis was performed using a one-way analysis of variance (ANOVA), and a probability $P < 0.05$ was considered as statistically significant.

3 RESULT

3.1 Removal of phenol by *I. galbana* under different temperatures

3.1.1 The effect of temperature on *I. galbana* growth and the removal of phenol

The attenuation curve of phenol by *I. galbana* is shown in Fig.1, and the corresponding curve of cell growth is shown in Fig.2. The most suitable temperature for the growth of *I. galbana* among the five temperatures was 25°C, and the removal efficiency of phenol was the highest at the same temperature. The growth of the microalga was inhibited with a decrease in temperature, and the removal efficiency of the substrate decreased accordingly. The high temperature (30°C) also slowed down the growth. The microalga nearly completely removed 100 mg/L of phenol in 24 h at 25°C and 20°C. The removal of phenol by *I. galbana* was inhibited at lower temperatures, such as 10°C.

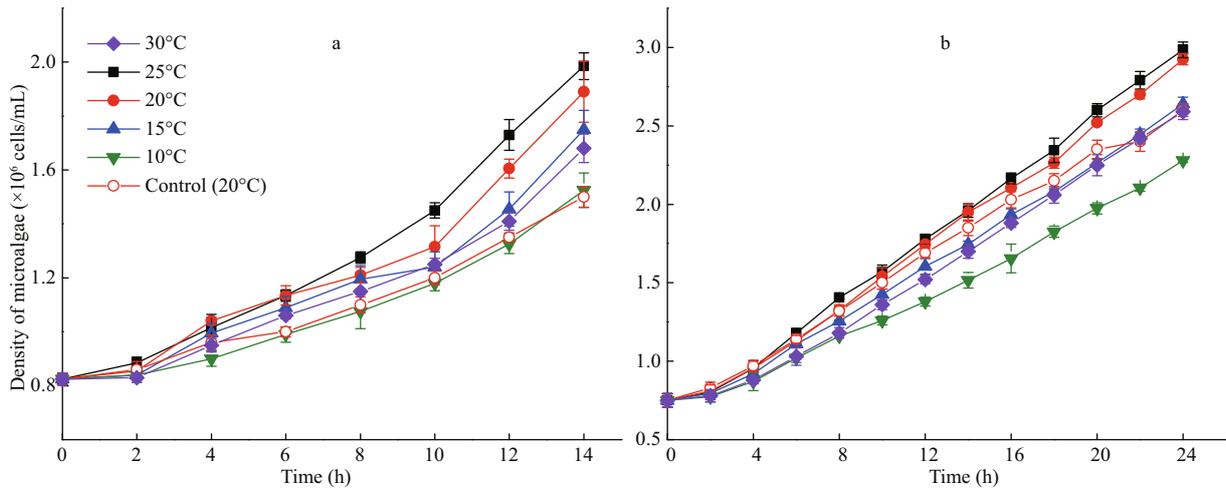


Fig.2 Growth curves of *I. galbana* exposed to 50 mg/L (a) and 100 mg/L (b) of phenol under different temperatures

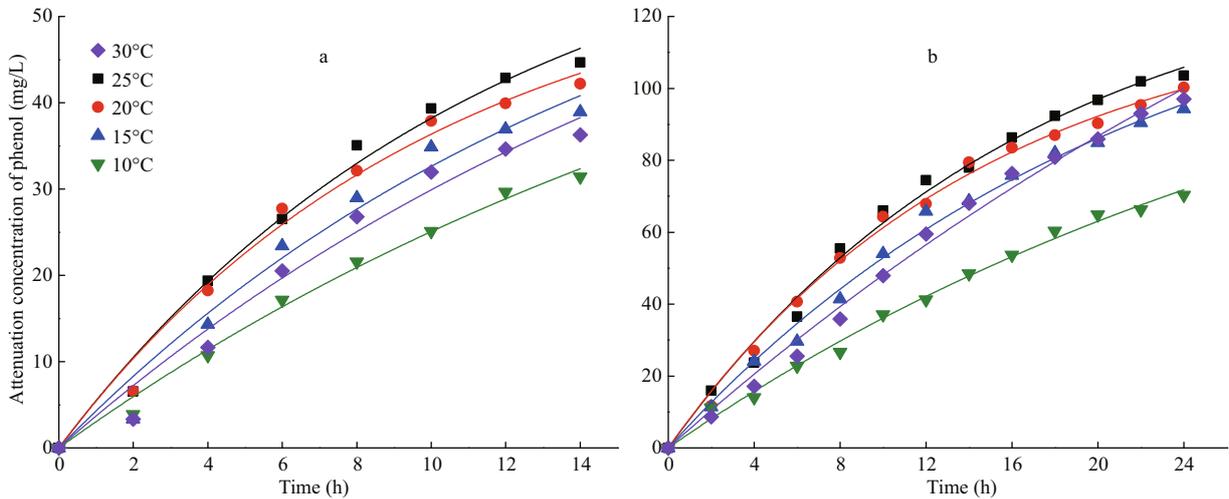


Fig.3 Pseudo first-order profiles for the biodegradation of phenol with 50 mg/L (a) and 100 mg/L (b) using *I. galbana* under different temperatures

3.1.2 Kinetics of phenol removal by *I. galbana* under different temperatures

The exponential rate model was used to characterize the attenuation process of phenol by *I. galbana* in this study. The aim was to compare the degradation effect under diverse temperatures using a constant degradation rate. Hence, zero-order and first-order kinetic models were applied to fit the degradation process of microalga at different temperatures, and the best-fit exponential rate model was chosen to describe the removal process of phenol. The fitting curves of the kinetics are shown in Fig.3, and the corresponding parameters are detailed in Table 1. The degradation kinetics of *I. galbana* in two concentrations of phenol both fit to the first-order model, and all of the R^2 values were higher than 0.972.

A comparison with the K values under different

Table 1 First-order kinetic constants and half-lives for the biodegradation of phenol by *I. galbana* under different temperatures

Phenol concentration	Temperature (°C)	K	Half-life (h)	R^2
50 mg/L	30	0.062 3±0.005 2	11.2±0.9	0.974
	25	0.087 7±0.025 0	8.3±2.3	0.986
	20	0.098 6±0.006 2	7.0±0.5	0.985
	15	0.068 1±0.009 3	10.3±1.4	0.972
	10	0.044 6±0.016 0	17.2±7.0	0.991
100 mg/L	30	0.021 9±0.000 1	31.7±0.1	0.992
	25	0.060 0±0.000 2	11.6±0.0	0.993
	20	0.068 1±0.003 9	10.2±0.6	0.996
	15	0.046 5±0.000 1	14.9±0.0	0.994
	10	0.029 2±0.017 7	29.0±17.5	0.994

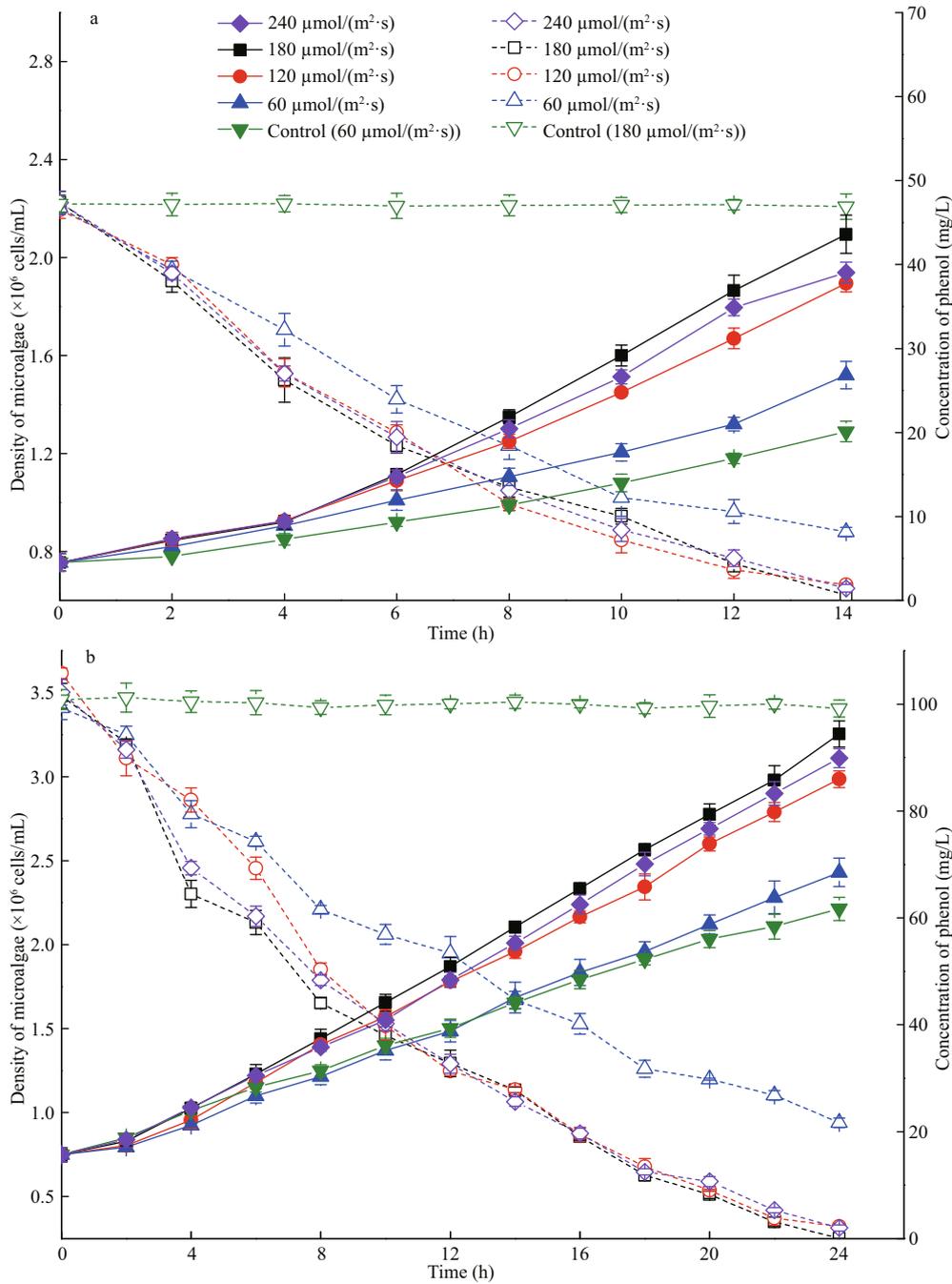


Fig.4 Phenol removal and cell growth of *I. galbana* in 50 mg/L (a) and 100 mg/L (b) of phenol under different light intensities
 Solid points represent the cell densities, and empty points represent the phenol concentrations. Control (60 $\mu\text{mol}/(\text{m}^2\cdot\text{s})$) represents the control without phenol, and control (180 $\mu\text{mol}/(\text{m}^2\cdot\text{s})$) represents the control without microorganism.

temperatures (Table 1) showed that the optimal temperatures of phenol removal for 50 and 100 mg/L of phenol were both 20°C, and the values of *K* were 0.098 6 and 0.068 1/h, respectively. In addition, the half-lives of phenol were 7.0 and 10.2 h. Although the results were consistent with those shown in Fig.1, the removal efficiency of phenol at 20°C was close to 25°C (Fig.1).

3.2 Removal of phenol by *I. galbana* under different light intensities

3.2.1 The effect of light intensity on *I. galbana* growth and the degradation of phenol

The attenuation curves of phenol and growth curves of microalga under different light intensities are shown in Fig.4. The curves of microalga in complete darkness

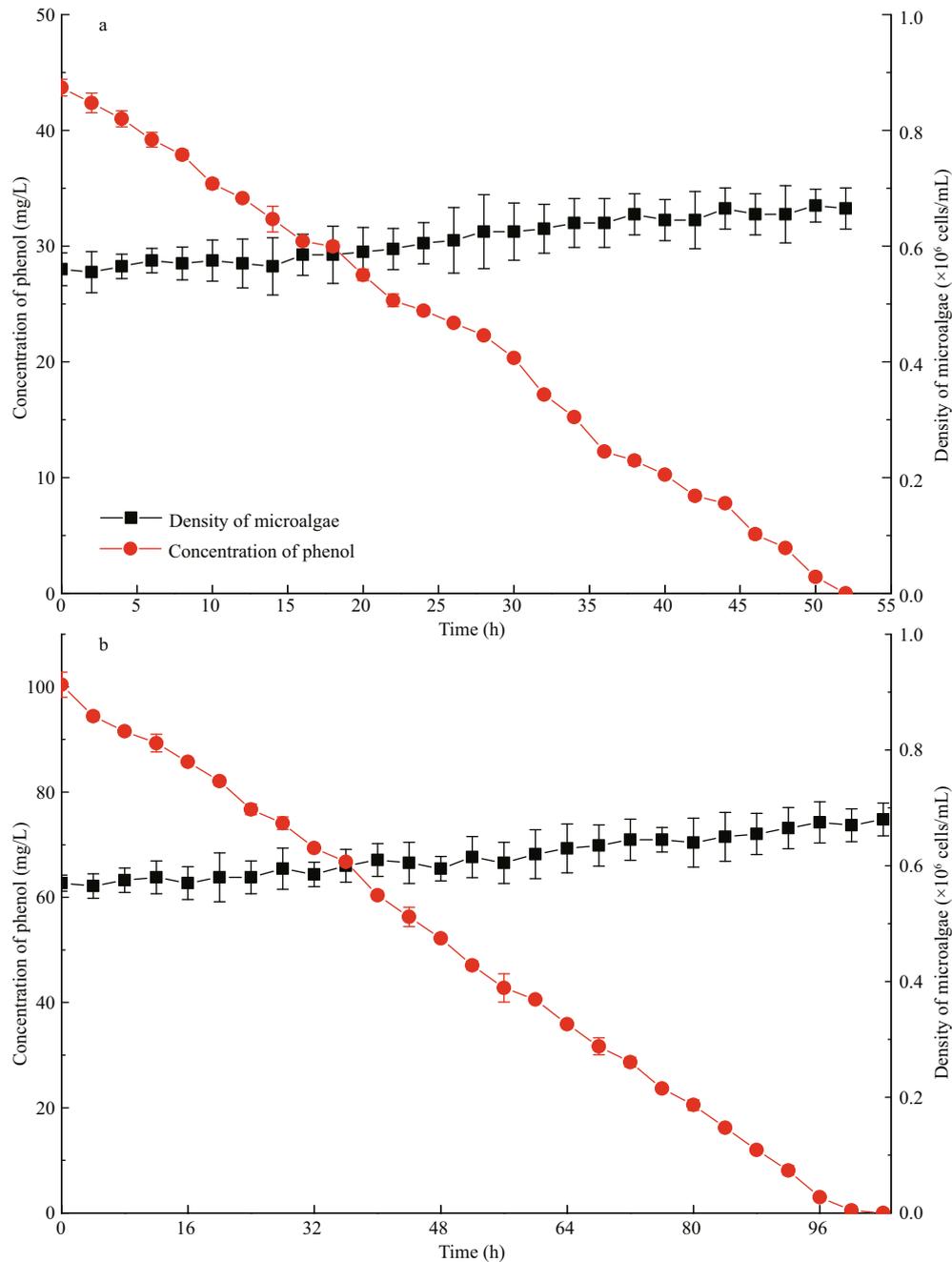


Fig.5 Removal in 50 mg/L (a) and 100 mg/L (b) of phenol with *I. galbana*, and cell growth under dark conditions

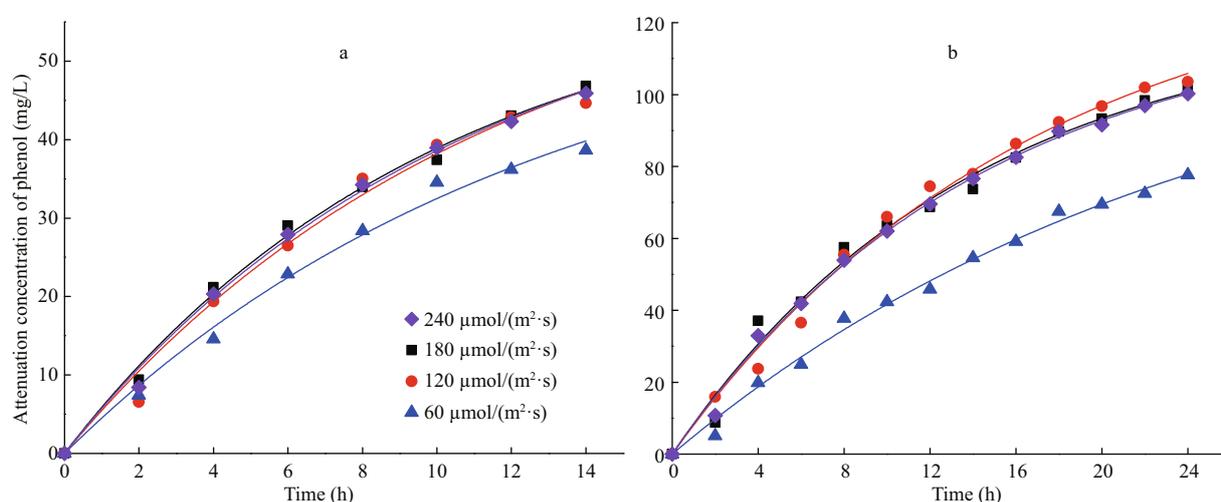
are shown in Fig.5. According to the growth curves, $180 \mu\text{mol}/(\text{m}^2 \cdot \text{s})$ was most suitable for the growth of *I. galbana* among the four light intensities. The growth of microalga was inhibited at a low illumination condition. However, the removal trend of phenol by microalga was different from the growth change at the various illuminations. The degradation of phenol under $180 \mu\text{mol}/(\text{m}^2 \cdot \text{s})$ was similar to the result of 120 and $240 \mu\text{mol}/(\text{m}^2 \cdot \text{s})$, and their reduction curves were almost coincident. The 50 and 100 mg/L of phenol were nearly completely removed by *I. galbana* in 14 and

24 h, respectively, under the three higher light intensities. However, the degradation effect of phenol and the growth of microalga at $60 \mu\text{mol}/(\text{m}^2 \cdot \text{s})$ were less satisfactory in comparison to the other light intensities.

Phenol could be removed by microalga under dark conditions (Fig.5), and light may be not an essential factor for the removal of phenol. However, 50 and 100 mg/L of phenol were completely removed in 52 and 102 h, respectively. Hence, the degradation time was much longer than the time under the light condition (14 and 24 h at $180 \mu\text{mol}/(\text{m}^2 \cdot \text{s})$).

Table 2 Calculated first-order kinetic constants and half-lives for the removal of phenol with *I. galbana* under different light intensities

Concentration of phenol	Light intensity ($\mu\text{mol}/(\text{m}^2\cdot\text{s})$)	K	Half-life (h)	R^2
50 mg/L	240	0.096 0 \pm 0.022 1 (First-order)	7.4 \pm 1.7	0.995
	180	0.100 9 \pm 0.009 7 (First-order)	6.9 \pm 0.6	0.995
	120	0.087 7 \pm 0.025 0 (First-order)	8.3 \pm 2.3	0.986
	60	0.078 4 \pm 0.014 1 (First-order)	9.0 \pm 1.7	0.991
	0 (dark)	0.848 7 \pm 0.013 1 (Zero-order)	25.7 \pm 0.0	0.997
100 mg/L	240	0.070 0 \pm 0.009 1 (First-order)	10.0 \pm 1.3	0.996
	180	0.072 0 \pm 0.000 2 (First-order)	9.6 \pm 0.0	0.988
	120	0.059 4 \pm 0.000 7 (First-order)	11.6 \pm 0.1	0.993
	60	0.040 4 \pm 0.008 2 (First-order)	17.2 \pm 3.5	0.993
	0 (dark)	1.001 8 \pm 0.016 0 (Zero-order)	50.1 \pm 0.4	0.998

**Fig.6** Pseudo first-order profiles for the removal of phenol with 50 mg/L (a) and 100 mg/L (b) using *I. galbana* under different light intensities

3.2.2 Kinetics of phenol removal by *I. galbana* under different light intensities

The kinetic fitting curves are shown in Fig.6, and the corresponding parameters are detailed in Table 2. The kinetics in two concentrations of phenol both fit to the first-order model under different light intensities, and all of the coefficients of determination (R^2) were higher than 0.986. The kinetics of phenol removal by *I. galbana* under dark conditions fit to the zero-order model.

As shown in Table 2, the most suitable light intensity for the removal by *I. galbana* was 180 $\mu\text{mol}/(\text{m}^2\cdot\text{s})$, the constants of the degradation rate (K) were 0.100 9 and 0.072 0/h at 50 and 100 mg/L of phenol, respectively, and half-lives were 6.9 and 9.6 h, respectively. The optimal light intensities for the microalgae growth and removal of phenol were the same.

4 DISCUSSION

Low concentrations of organic pollutants may promote the growth of microorganisms. According to Figs.2 & 4, 50 mg/L and 100 mg/L of phenol can promote the growth of *I. galbana* compare with control experiments. However, Wang et al. (2019) indicated a phenol concentration ≥ 125 mg/L inhibits the reproduction of *I. galbana*. Di Caprio et al. (2018) drew a similar result that phenols concentration higher than 100 mg/L have an inhibition effect on *Scenedesmus* sp.

The optimal temperatures for removal of phenol and growth of microalgae are 25°C and 20°C respectively through experiments and kinetic model. Considering the removal effects of phenol at 25°C and 20°C were close (Fig.1 and Table 1), the temperature range between 20°C to 25°C can be seen

as appropriate temperatures for the removal of phenol by *I. galbana*. But the removal of phenol is inhibited by low temperatures, the reason may be that lower temperatures inhibited the activities of enzymes in the microbes so that their growth and metabolism of microorganisms were influenced (Siddiqui, 2015). Higher temperatures improve the activities of the enzymes related to photosynthesis, and relevant processes, such as the diffusion of CO₂, were also boosted (Raven and Geider, 1988). In addition, higher temperatures may speed up the process of cellular metabolism to promote the growth of microalga. When the temperature continues to increase and exceed the optimum temperature, irreversible physiological reactions may happen in cells, such as a decrease in the photosynthetic rate and growth rates of microorganisms. Also, the alga may be susceptible to other environmental factors, such as irradiance and salinity (Wernberg et al., 2010; Andersen et al., 2013). So the growth of *I. galbana* and phenol removal became to be inhibited at 30°C. Thus, it can be seen that the removal process of phenol by *I. galbana* is affected by temperature of the marine environment.

The effect of temperature on the removal of phenol with microorganisms has been studied in many reports. Among them, studies of bacteria have been the most prevalent. Onysko et al. (2000) found that the optimum temperature for biodegradation of phenol by *Pseudomonas putida* was 25°C, and the kinetic parameters of degradation showed an increasing trend with an increase in temperature. Polymenakou and Stephanou (2005) studied the removal effect of phenol by a native *Pseudomonas* sp. between 10°C and 40°C and confirmed 30°C was best for the degradation of phenol for *Pseudomonas* sp. Lu et al. (2009) studied the degradation of phenol by *Phanerochaete chrysosporium* at different temperatures, and the results showed that first-order kinetics were suitable for the biodegradation process, and the optimal temperature was 37°C for the removal of phenol. In comparison with these temperatures, the most suitable temperature of *I. galbana* (20–25°C) is relatively low, and this temperature can be more easily achieved in actual marine environments.

Wang et al. (2017a) studied the natural attenuation of phenol under simulated marine conditions, the results indicated that photolysis was not the main pathway of phenol removal. And according to Fig.4, the phenol concentration of control groups (containing 50 mg/L and 100 mg/L of phenol without microalga) did not change in the whole process. Therefore, the

effect of photolysis is neglectable in this study. As shown in Fig.4, the best effect of growth of *I. galbana* and removal of phenol is obtained at 180 μmol/(m²·s) (Fig.4), and the kinetic model agree with the results (Table 2). However, in a marine environment, the intensity of light decreases with an increase in the water depth, and the irradiance in deeper waters is generally very weak (Wang et al., 2017b). By inference, the degradation efficiency of phenol by *I. galbana* will decrease as water depth increases. This phenomenon indicates that the removal process progressed slowly instead of terminating in a dark environment. At the scene of an actual marine accident, the dark or weak light conditions will slow down the process of phenol removal by microalga, instead of completely inhibiting it. Moreover, the growth of the microalga under dark conditions was extremely inhibited, and there was very little growth in the cell densities of the microalga at 96 h (Fig.5) because the growth of *I. galbana* was inhibited in heterotrophy (Alkhamis and Qin, 2013). It is speculated that *I. galbana* could remove phenol under dark conditions, but phenol would not be utilized as the carbon source for the growth of this microbe. Nazos et al. (2017) studied the biodegradation of phenol by *Chlamydomonas reinhardtii* under different light intensities. Their results showed that degradation was a process of energy consumption. The metabolism of phenol by *C. reinhardtii* only occurred under aerobic and light conditions. Hence, the phenol was not degraded during anaerobic or dark conditions. This conclusion is different from the results of this study, and this may be due to the different microalgae species.

The removal mechanism of organic pollutants by microalgae mainly includes biosorption, biodegradation and bioaccumulation. According to Song et al. (2019) and Xiong et al. (2017), biodegradation is the main pathway on the removal of organic contaminants while biosorption and bioaccumulation have little contribution to the removal process. However, the mechanism of phenol removal by *I. galbana* will be studied in the future. Furthermore, most microalgae inhabit the surface water due to their phototaxis. Therefore, in situ bioremediation via microalgae is suitable for the removal of phenol pollution in shallow water and coastal seawater. New biotechnologies, such as immobilization, are in urgent need for treating phenol contaminations in deeper seawater. In addition, microalgae is expected to be used in wastewater

treatment, e.g. a *Leptolyngbya* sp. strain was applied to efficiently remove 100 mg/L of phenol from coke-oven wastewater (Thakurta et al., 2018). Hence *I. galbana* can also be used to the ex situ treatment of saline phenol-containing wastewater.

5 CONCLUSION

The effects of temperature and light intensity on the removal of phenol by *I. galbana* in seawater were experimentally studied in this research. The results showed that the kinetics under different temperatures and light intensities agreed with the first-order model. Temperature and illumination have the ability to significantly influence the bioremediation effect. The most suitable temperature and light intensity were 20°C and 180 $\mu\text{mol}/(\text{m}^2\cdot\text{s})$, respectively. Phenol was slowly removed by *I. galbana* in dark conditions, and light was not essential for the removal of *I. galbana*, but the microalga did not grow under dark conditions. Although the inhibition could occur, *I. galbana* was shown in this study to remove high concentrations of phenol at low-temperatures or under low-light (or dark) conditions. Hence, *I. galbana* has a promising potential for applications in the bioremediation of phenol pollution in actual marine environments.

6 DADA AVAILABILITY STATEMENT

The data generated or analyzed during the current study are available from the corresponding author on reasonable request.

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